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Household water treatment and safe storage – effectiveness and economics

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Abstract

Household Water Treatment and safe Storage (HWTS) systems aim to provide safe drinking water in an affordable manner to users where safe piped water supply is either not feasible or not reliable. In this study the effectiveness, costs and cost drivers

- ⁵ of three selected HWTS systems were identified. The selected systems are SODIS, ceramic filter and biosand filter. These options were selected based on their current usage rate, available scientific data, and future potential. Data was obtained through peer-reviewed literature, reports, web-pages and informal sources. The findings show a wide dispersion for log removal of effectiveness of the HWTS systems. For bacte-
- ria, log removals of 1–9 (SODIS), 0.5–7.2 (ceramic) and 0–3 (biosand) were reported. In the case of viruses, log removals of 0–4.3 (SODIS), 0.09–2.4 (ceramic) and 0–7 (biosand) were found. The dispersions of log removal for both bacteria and viruses range from non-protective to highly protective according to WHO performance targets. The reported costs of HWTS systems show a wide range as well. The price per cubic
- ¹⁵ meter water is found to be EUR 0–8 (SODIS), EUR 0.37–6.4 (ceramic) and EUR 0.08– 12.3 (biosand). The retail prices found are: negligible (SODIS), USD 1.9–30 (ceramic) and USD 7–100 (biosand). No relationship was observed between removal efficiency and economics of the three systems.

1 Introduction

In many parts of the world, people do not have access to safe drinking water, this is especially true in rural areas of developing countries (Unicef et al., 2012). Conventional piped water or similar centralized systems are decades away for these people and they are often left with the responsibility and need to collect, treat and store their own water (Brown et al., 2008). Where groundwater is inaccessible or contaminated, these users
 depend on household water treatment (HWTS) systems for safe drinking water (Sobsey et al., 2008). These HWTS systems have the goal to provide safe drinking water in an





affordable and sustainable manner (Duke et al., 2006) while being simple and easy to manage by their users (Heinsbroek and Peters, 2014). As such, these systems are crucial in reducing occurrence of diarrheal and other debilitating illnesses (Meierhofer and Landolt, 2009; Stauber et al., 2009). Efficiency in providing safe water differs per method. To indicate removal efficiency, the WHO produced guidelines (WHO, 2004) to define default performance targets to indicate a certain removal efficiency for different pathogens as "interim", "protective" or "highly protective" (see Fig. 1).

 $Price [EURm^{-3}] = \frac{Investment + operational costs}{Produced water}$ (1)

When looking at the economics of HWTS systems, it is common practise to look at
 the price per produced m³ water (NWP, 2010). Generally, this is calculated by dividing the investment and operational costs over the produced water during the lifetime of the technology (NWP, 2010).

The objective of this paper is to give an overview of the potential effectiveness according to the WHO performance targets and the costs paid by the user of three HWTS

systems: SODIS, ceramic filters and biosand filters. These HWTS systems are selected because they are widely used, they are promising for the future and sufficient academic literature is written to research their removal efficiency.

1.1 SODIS

SODIS is based on the principle of disinfection by solar radiation (see Fig. 2). The procedure is extremely simple; an unscratched and uncoloured PET or glass bottle is filled with water and exposed to direct sunlight for a minimum of 6 h (Heinsbroek and Peters, 2014). Water with low oxygen and high turbidity levels has to be pre-treated (Acra et al., 1990; Meierhofer and Landolt, 2009).



1.2 Ceramic filters

The ceramic filter is based on the following principle: a porous media of fired clay that removes microbes by size exclusion and tortuosity (ratio of shortest linear distance over actual flow path) (Hunter, 2009; Sobsey et al., 2008; van der Laan et al., 2014).

⁵ Many variations of ceramic filters exist; e.g. pot filters or "water purifier" (see Fig. 3) (Akvopedia, 2014b; Potters for Peace, 2014), candle filters (Sobsey et al., 2008) and Tulip siphon filter (Basic Water Needs, 2014; Tulipfilter, 2013). Periodic scrubbing and rinsing is necessary to remove impurities (Sobsey et al., 2008).

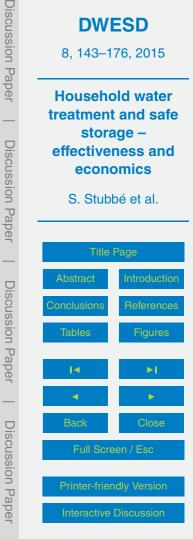
1.3 Biosand filter

- ¹⁰ Biosand filters consist of a concrete or plastic frame filled with crushed rock (sand) filter media of 0.15–0.35 mm particle (Murphy et al., 2010b) (see Fig. 4). Two filter mechanisms govern the removal principle of biosand filters: physical removal of organic matter and turbidity (Sobsey et al., 2008) and biological removal of colloidal particles and harmful pathogens in the so-called Schmutzdeke (Duke et al., 2006; Hunter, 2009;
 ¹⁵ Weber-Shirk and Dick, 1997). The filter can be cleaned manually by removing the top
- few centimetres of sand and disposing the overlying water (Sobsey et al., 2008).

2 Effectiveness

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In this section an in-depth description is given of the removal mechanisms of each of the selected HWTS systems and the corresponding removal efficiency. Both lab and field studies are used to give an overview of the reported effectiveness. Insufficient data was found on the log removal of protozoa, so this pathogen is excluded from this study.





2.1 SODIS

The inactivation mechanisms of the solar radiation is based on direct UVB absorption (damaging the pathogenic DNA), optical inactivation (via reactive oxygen species) and thermal inactivation (denaturation) (Reed, 2004). A synergy between optical inactivation and thermal inactivation was signalled at temperatures between 40.50°C

activation and thermal inactivation was signalled at temperatures between 40–50 °C (Reed, 2004; Wegelin et al., 1994). Several parameters are suggested to enhance the SODIS treatment: black background surface (to reflect sunlight) (Martín-Domínguez et al., 2005; Wegelin and Sommer, 1998), unscratched container material (diminish scattering) (Wegelin and De Stoop, 1999), added photosensitizers (increase production of oxygen reactive species) (Chilvers et al., 1999) and glass bottles (Duffy et al., 2004). Critics are focused on the potential leaching of plasticizers into the treated was the (De Stoop) and the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of plasticizers into the treated was an end of the potential leaching of the potential leaching of plasticizers into the potential leaching

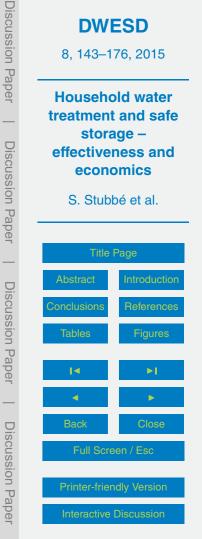
ter (Reed, 2004). However, Wegelin (2001) has shown that this is only the case at the outer surface of the bottles and not in the treated water.

In Figs. 5 and 6, a summary is given of the found removal efficiencies of SODIS for

- ¹⁵ bacteria and viruses respectively. The majority of the research conditions lays between 40–65 °C and 4–6 h. For bacteria and viruses the log removal was between 1–9 and 0–4.3 respectively based on: Acra et al. (1990); Akvopedia (2013); Dejung et al. (2007); Fujioka and Yoneyama (2002); Heaselgrave et al. (2006); Joyce et al. (1996); Lonnen (2005); Martín-Domínguez et al. (2005); McGuigan et al. (2012); Meyer and Reed (2004); Sodis.ch (2011). The majority of the results are centred around 2.5–5
- log removal and 1-4 log removal for bacteria and viruses respectively.

2.2 Ceramic filters

By means of meta-regression, Hunter (2009) concluded that compared to other interventions (chlorine, SODIS, biosand filter and combined coagulant-chlorine), the ce-²⁵ ramic filter shows the highest effectiveness on the long term. Most filters are manufactured by adding colloidal silver to increase efficiency. Silver inactivates bacteria and other pathogens through three mechanisms: reaction with thiol (in structural groups



and functional proteins), structural changes in cell membrane and reaction with nucleic acids (Russell et al., 1994). There are different ways to impregnate silver in the filter: dipping, painting, pulse injections and fire-in (Oyanedel-Craver and Smith, 2007; Ren and Smith, 2013). Van der Laan (2013) and Oyanedel (2008) did not find a significant
⁵ difference in removal efficiency for different silver application methods. Neither does an addition of iron appear to increase the removal efficiency in their research (Brown et al., 2008). On the contrary, the storage time in the receptacle of the filter was found to be an important parameter in the bacterial removal efficiency (van der Laan et al., 2014). Concerns exist about the virus removal of ceramic filters, since reported removal efficiencies do not reach WHO guidelines (Murphy et al., 2010a; van der Laan et al., 2014), and show high dispersion (Bielefeldt et al., 2010). No critical parameter was yet identified to improve the virus removal efficiency (van der Laan et al., 2014).

In Figs. 7 and 8, a summary is given of the found removal efficiencies of ceramic filters for bacteria and viruses respectively. The log removal of ceramic filters for bacte-

- ria and viruses were between 0.5–7 and 0.09–2.4 respectively based on: Basic Water Needs (2014); Bielefeldt et al. (2010); Bloem et al. (2009); Brown et al. (2008); Brown and Sobsey (2010); Lantagne (2001); Murphy et al. (2010a); Oyanedel-Craver and Smith (2007); Potters for Peace (2014); Roberts (2003); Simonis and Basson (2011); Tulipfilter (2013); van der Laan et al. (2014); Van Halem et al. (2007). For removal of bacteria, most information sources report log removals between 1.3 and 4. The log
- removals of viruses are significantly lower with most information sources reporting log removals between 0.4 and 1.4.

2.3 Biosand filters

The removal mechanism of the biosand filter is based on the slow sand filtration principle and depends on the daily volume charged to the filter (Elliott et al., 2008). The optimal charge volume is investigated to be equal or smaller than the pore volume (Elliott et al., 2011). When larger charge volumes are exposed to the filter, a decrease in removal efficiency is found (Baumgartner et al., 2007). Although this HWTS system is



designed for intermitted use, research shows that continuous use of the biosand filter has higher removal efficiencies (Young-Rojanschi and Madramootoo, 2014). Introduction of iron oxide in the sand layer shows improved levels of pathogen removal and is especially beneficial after cleaning or in the ripening period (Ahammed and Davra,

- ⁵ 2011). The mechanism behind this enhanced microbial removal is lacking a clear explanation, but possible explanations are: hydrophobic interactions, macromolecular bridging, electrical double-layer and van der Waals forces (Kim et al., 2008; Truesdail et al., 1998). It is suggested that the Schmutzdeke contributes to the virus attenuation by the production of microbial exo-products (proteolytic enzymes) or grazing bacteria on
- virus particles (Elliott et al., 2011; Huisman et al., 1974). Concerns exist about the nonexisting guidelines for the post-treatment of the removed Schmutzdeke during maintenance since this contains opportunistic pathogens and therefore poses an health risk to consumers (Hwang et al., 2014).

Figures 9 and 10 provide a summary of the reported removal efficiencies of biosand
filters. Overall, the reported log removals of biosand filters for bacteria and viruses are between 0–3 and 0–7, respectively (based on: Ahammed and Davra, 2011; Akvopedia, 2014c; Baumgartner et al., 2007; Duke et al., 2006; Elliott et al., 2008; Murphy et al., 2010a; Palmateer, 1999; Sswm.info, 2014; Stauber et al., 2009; Vanderzwaag et al., 2009; Wang et al., 2014). The log removal of bacteria is centred on 0.4–2; while the distribution of log removals reported for viruses is widely scattered.

2.4 Overview of effectiveness

Figures 5–10 show that the removal efficiency of HWTS systems differs per pathogen type and per research study. Peer-reviewed research is in this case more reliable than other studies, but even these results do not always agree. The removal efficiencies found in the reviewed articles, are not always compatible with the target performance of the WHO (see Fig. 1), which corresponds with the results of previous studies such as (Murphy et al., 2010a; van der Laan et al., 2014). The difference between highest and lowest reported efficiencies of each HWTS system is exactly what makes the dif-





ference between safe or unsafe water produced with this filter. Hence, the question arises whether certain removal efficiency can be guaranteed for the HWTS systems.

In Figs. 11 and 12, the total range of lowest to highest log removal reported is shown per HWTS system. It is assumed that every reported removal efficiency has the same

⁵ likelihood to occur. It can be seen that SODIS have the highest reported efficiency for bacteria removal (9 log removal) whereas, biosand filter report the highest reported efficiency for virus removal (7 log removal). Biosand filters show the lowest (zero) removal efficiency for bacteria whereas for virus removal, all filters have been reported with a zero log removal in one or more studies.

10 2.5 Critical evaluation of effectiveness

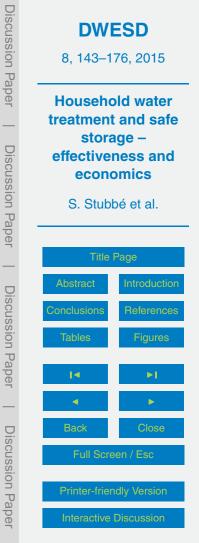
Recent studies put critical notes to the effectiveness of HWTS systems. Results of field tests reveal that HWTS systems do not always improve and sometimes even worsen the pathogenic state of the influent (Murphy et al., 2010a). The lack of blinding and considerable heterogeneity in the results of HWTS systems show signs of concerns (Hunter, 2009). Moreover, it is reported that the research method can have a big impact

¹⁵ (Hunter, 2009). Moreover, it is reported that the research method can have a big impact on the reported efficiency (van der Laan et al., 2014). The reported removal efficiency also depends on the indicator pathogen used, as shown by Palmateer (1999) and Elliot (2008). Quality tests are not yet globally standardized (Rayner et al., 2013), so that a fair comparison between data sets is challenging.

20 2.6 Human factors

Research shows that operating conditions can reduce the effectiveness of HWTS systems (Baumgartner et al., 2007). The effectiveness of HWTS systems does not only depend on technology, but also on human factors. When the HWTS system is not operated properly, exposure to pathogens can remain high. For example, it is common

that people use the storage container of the device to collect dirty untreated water to feed the HWTS system, reducing the effectiveness of the device (Murphy et al.,





2010a). Other reasons why in practice the effectiveness of HWTS systems is reduced:
(i) people only treat part of their used water (Sobsey et al., 2008), as the water supply of HWTS systems can be reduced in time due to clogging (ii) people are unable to purchase replacements (Brown et al., 2009; Hunter, 2009; Meierhofer and Landolt, 2009), (iii) people only treat their water intermittently (Sobsey et al., 2008), (iv) people have limited guidance to determine whether pre-treatment is pecessary (Sobsey et al.)

- have limited guidance to determine whether pre-treatment is necessary (Sobsey et al., 2008), (v) people have simply stopped using the device (Hunter, 2009), (vi) or have sold it to a friend or relative (Brown et al., 2009). For ceramic filters, the rate of participation reduction is estimated at 2 % month⁻¹ (Brown et al., 2009). The found diversity
- in effectiveness prompts that sufficient training and continued monitoring is needed to increase and sustain proper HWTS device management. Preferably, this could be done by a well-embedded local agent in order to increase acceptability (Meierhofer, 2006). Understanding the human factors that influence the real effectiveness of the HWTS systems is crucial for widespread adoption and sustained usage (Sobsey et al., 2008).

3 Economical evaluation

In this section, the parameters that determine the purchase price of a HWTS system (see Fig. 12) and the reported prices for the three selected HWTS systems are discussed (see Table 1).

3.1 Economic Parameters

- The price of HWTS systems depends strongly on project area (K. Wagoner, personal communication, 2014; H. Jansen, personal communication, 2014). This can be explained by the fact that the price of HWTS systems is determined by (at least) 5 parameters (see Fig. 12). The first parameter of the costs for HWTS systems is the production costs (or investment costs) (Basic Water Needs, personal communication, 2014). The time approximation (2014). These posts depend on the time (2014).
- ²⁵ 2014; K. Wagoner, personal communication, 2014). These costs depend on the type



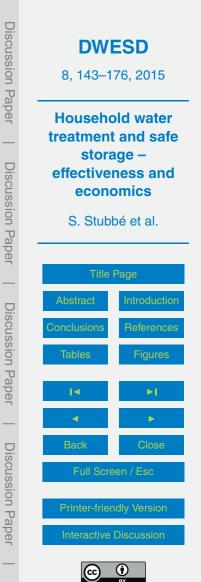


of HWTS system and the region of production. Factories in China and India are frequently used, due to lower labour costs. The second parameter is distribution (Stuurman et al., 2010). Transport costs between production and project area depend on quantity and weight. In-land and over-sea transportation can differ significantly in to-

- tal cost (Basic Water Needs, personal communication, 2014). This parameter is estimated to be the most dominant (Basic Water Needs, personal communication, 2014). However, transportation costs are often not included in the reported price values on websites or in literature, because of the high variability per project area. Local production factories are established to diminish distribution costs and enhance local economy
- (Brown, 2007). The third parameter is taxes. Depending on the country, HWTS systems need to be imported and import fees are involved. These costs can be relatively high (Basic Water Needs, personal communication, 2014). A possible fourth parameter is whether subsidies are present for the project (Stuurman et al., 2010). Often, NGOs arrange donor projects where consumers pay a reduced price (Stuurman et al., 2010).
- ¹⁵ However, research points out that consumers use the provided filters more, when they really have to invest to purchase the technology (Brown et al., 2009). A possible fifth parameter is the (local) distributor's marginal fee needed to maintain the business (Basic Water Needs, personal communication, 2014). Depending how the supply chain is constructed, a (local) distributor organizes the sales in the project area. Furthermore,
 ²⁰ it should be mentioned that the price is time-dependent and susceptible to exchange
- ²⁰ It should be mentioned that the price is time-dependent and susceptible to exchang rates.

3.2 Costs of HWTS systems

A limited number of peer-reviewed articles mention costs of HWTS systems, and only retail prices were mentioned. Retail price depend on the 5 parameters mentioned in the previous section and is the price eventually paid by the user. The retail price could be converted to the price per m³, when the potential amount of water that can be treated with one filter is known (see Fig. 13). This potential amount is dependent on the lifetime of the filter, the flowrate, the sustainability of the system, e.g. Reliable and



sufficient information about this total potential amount is lacking. Therefore the retail price is mentioned separately from the price per m_1^3 water produced (see Table 1).

In Table 1, a summary is given of the price per m³ and retail price per HWTS system. SODIS has a retail price of USD0, since old PET bottles can be used. When new

- ⁵ bottles are used, only a small investment is necessary (NWP, 2010). The costs per m³ are related with the retail price. The outlier of 3–8 m⁻³ by (Akvopedia, 2013) is assumed to be an error, since it does not correspond with the numbers in the rest of the text of the same reference. For ceramic filters, the range of retail prices is between USD 1.9–30, with most of the references mentioning prices around USD 15. The differences in
- ¹⁰ price can be explained by the parameters elaborated in the previous section. The price per m³ ranges between USD 0.3 and USD 5.2, which depends on the estimation of the potential amount of water that can be treated with the filter. For biosand filters, large ranges are found in the price per m³ and retail price: EUR 0.07–10 and USD 7–100 respectively. The outlier in price per m³ of USD 10 is unreliable, since no argumentation is given in the reference (Akvopedia, 2014c). The outlier in retail price of USD 100 for
- concrete is also stated without further explanation (Sobsey et al., 2008).

Overall, it is found that the biosand filter has the lowest price per m³ produced what can be explained by its long life time, low maintenance costs and sustainable flowrate. Biosand filters do have the highest retail price (even after the highest outliner is ne-

²⁰ glected). Although SODIS is also a cheap technology, it requires a (small) investment when new (glass) bottles are used and it only produces little amount of water per bottle. SODIS does have the cheapest retail price. It is shown that ceramic filters have the biggest range of price per m³ water with the highest numbers. Ceramic filters are prone to breakage and the flowrate can decrease over time due to clogging. By far most independent research exists on ceramic filters compared to the other HWTS systems.

In Figs. 14 and 15, an overview of the price ranges is given, neglecting the outliners mentioned above.





Constraints to economic evaluation 3.3

Most cost estimations of HWTS systems are found on websites of coordinating NGOs or device suppliers. Because the information is practice-oriented, the reliability of this information is likely to be fluctuating. More direct information from local producers turned out to be necessary. Resource Development International in Cambodia reveals

a standing quotation of a ceramic filter for USD 12 (RDI, 2014), which is in line with the prices in other sources. Since the price of HWTS systems does not only depend on the 5 parameters mentioned before, but is also fluctuating in time and susceptible to exchange rates. The price of the HWTS system today is different from the price indicated for 2007 (H. Jansen, personal communication, 2014). This study does not include these changes. The prices mentioned in Table 1 are considered to be valid for the year of the respective reference.

Conclusions 4

In this study the economics and removal efficiencies of three selected Household Water Treatment and safe Storage (HWTS) systems were compared: SODIS, ceramic 15 filters and biosand filters. The costs of HWTS were based on five parameters: production, distribution, taxes, subsidies and marginal fees. They influence the price paid by the consumer besides other factors (interest, inflation). Additionally, the produced volume of water, or lifetime of the HWTS determines the actual price per m³. The re-

ported log removal should be viewed with some precaution, as parameters like indicator 20 pathogen, research method and human factors are of influence. Also, most studies are short-time (around 26 weeks) and designed poorly unblended, which could have given biased results (Hunter, 2009; Sobsey et al., 2008).

For SODIS, low retail prices and intermediate prices per m³ were observed with a range of removal efficiencies for bacteria from "non-protective" to "highly protective"

Discussion **DWESD** 8, 143–176, 2015 Paper Household water treatment and safe storage -Discussion effectiveness and economics S. Stubbé et al. Paper Abstract Discussion Conclusion Reference Tables Figures Paper Back Close Discussion Printer-friendly Versior Paper Interactive Discussion and for viruses from "non-protective" to "protective" according to WHO targets. Ceramic



filters showed intermediate retail prices and high prices per m³ with a range of removal efficiencies for bacteria from "non-protective" to "highly protective" and for viruses "non-protective". Biosand filters had high retail prices and low prices per m³ with a range of removal efficiencies for bacteria from "non-protective" to "protective" and for viruses from "non-protective" to "highly protective". Overall, a relationship between HWTS removal efficiency and economics was not observed.

The Supplement related to this article is available online at doi:10.5194/dwesd-8-143-2015-supplement.

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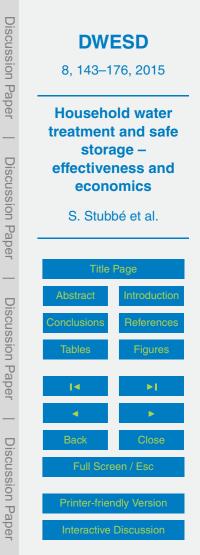
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Discussion

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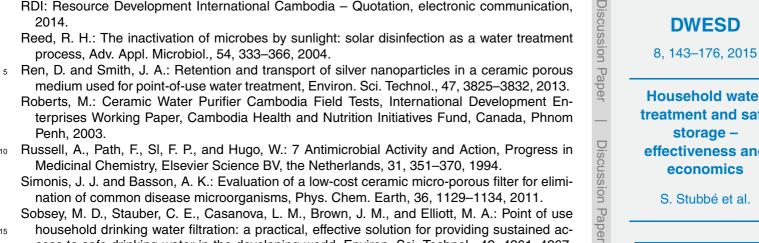
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Technology	Adjustment	Price (USDm ⁻³)*	Retail Price (USD)*	Reference
SODIS	_	0.53-2.03		NWP (2010)
		new	0	NWP (2010)
		0–0.27 used		akvopedia (2013)
		3–8	0	CAWST (2012)
Ceramic	General	0.3–0.61		NWP (2010)
			5–15	Oyanedel-Craver and Smith (2007)
			8–10	Brown et al. (2009), Sobsey et al. (2008)
	Tulip Siphon Filter	0.4-0.44	5.69-7.32	NWP (2010)
		1.14–5.2	26.02	Tulipfilter (2013)
		2–5		Basic Water Needs (2014)
	Pot Filter	1		akvopedia (2014b)
			9.76–17.89	CAWST (2012)
			7.5	Roberts (2003)
			5.4–28	NWP (2010)
	Water4life	0.46		akvopedia (2014a)
			1.87–20.33	NWP (2010)
	Candle Filter		15–30	CAWST (2012)
	Potters for Peace		15–25	CAWST (2012), Potters for Peace (2014b)
			12	(RDI, 2014)
Biosand	Concrete	0.07–0.15	6.99–22.76	NWP (2010)
		10	12–40	akvopedia (2014c), Sswm.info (2014)
			25–100	Sobsey et al. (2008), Ahammed and Davra (2011)
			7–28	CAWST (2012)
			12–30	CAWST (2012)
	Plastic		75	Ahammed and Davra (2011)
	Iron oxide filter		15–36	

Table 1. Overview of the price and retail price of HWTS systems.

* Conversion used where necessary 1.23 EUR/USD (Bloomberg, 2014).



Target	Log ₁₀ reduction required: Bacteria	Log ₁₀ reduction required: Viruses	Log ₁₀ reduction required: Protozoa
Highly protective	z 4	z 5	≥ 4
Protective	≥ 2	≥ 3	≥ 2
Interim*	Achieves "protective" target for two classes of pathogens and results in health gains		

Figure 1. WHO guidelines on default performance targets of HWTS systems (WHO, 2004).

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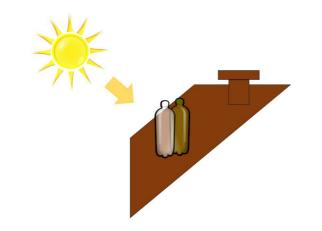
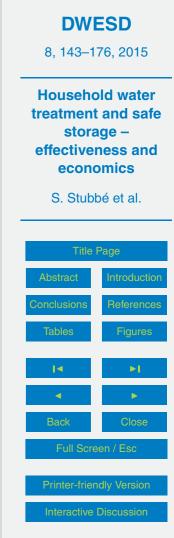


Figure 2. Schematic illustration of SODIS.



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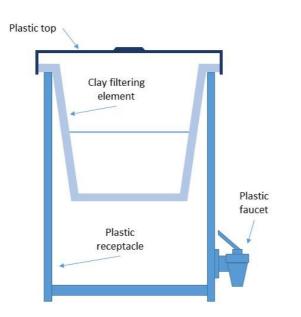
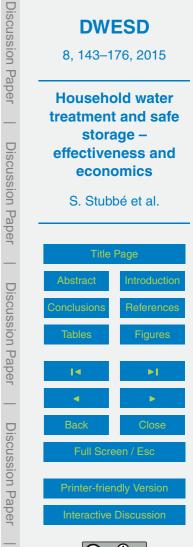


Figure 3. Schematic illustration of ceramic Pot filter (Van Halem et al., 2007).





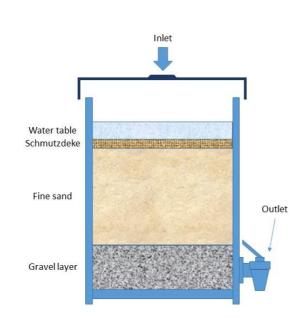
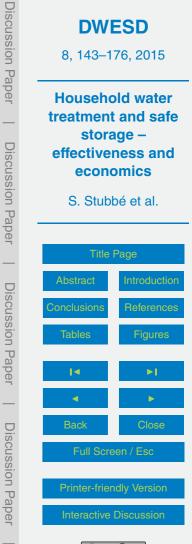
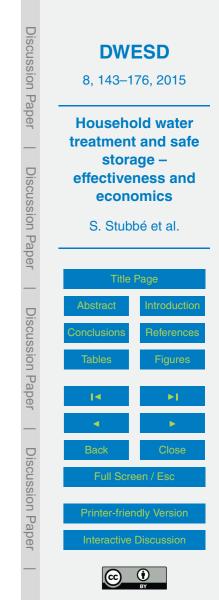


Figure 4. Schematic illustration of biosand filter.







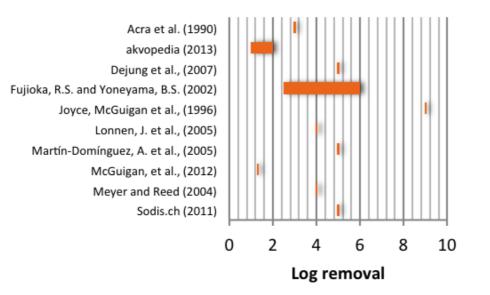
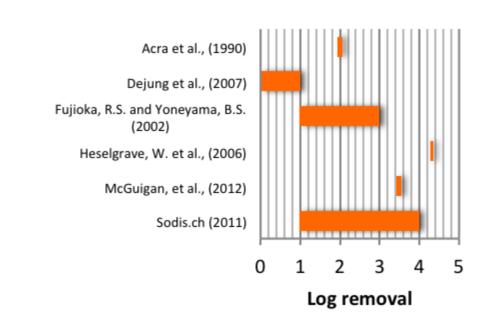
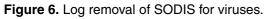
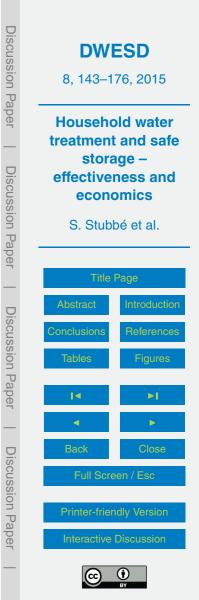


Figure 5. Log removal of SODIS for bacteria.







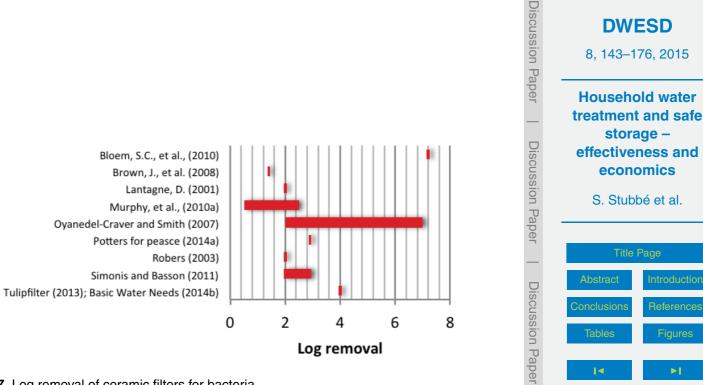
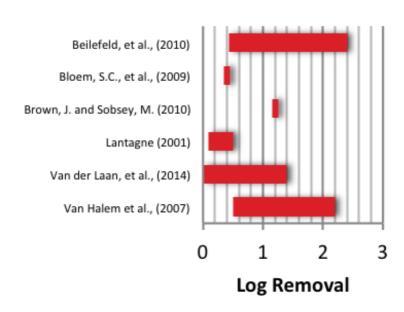
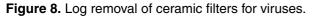


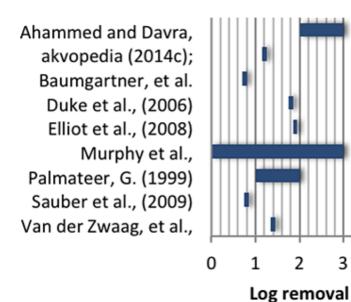
Figure 7. Log removal of ceramic filters for bacteria.

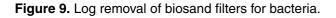


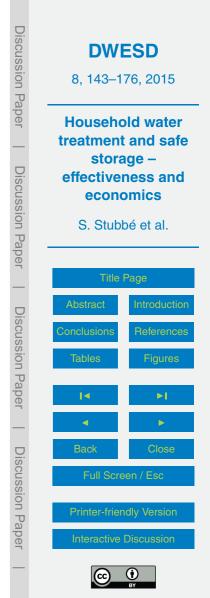


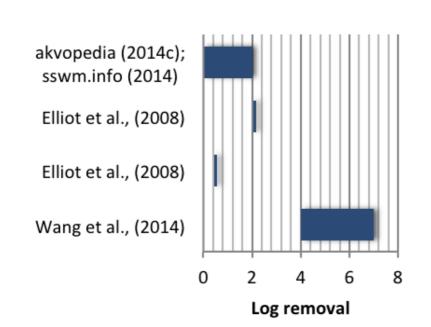


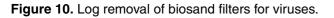


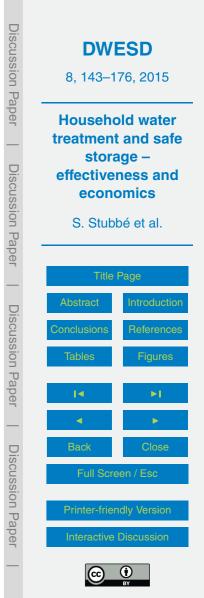


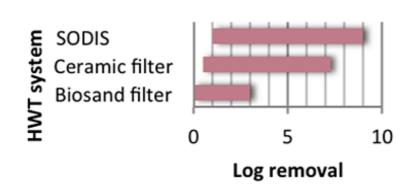


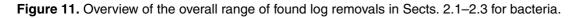




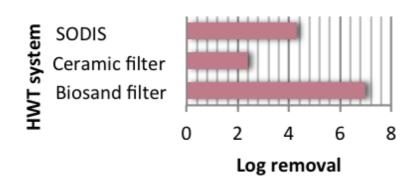


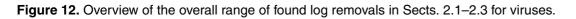






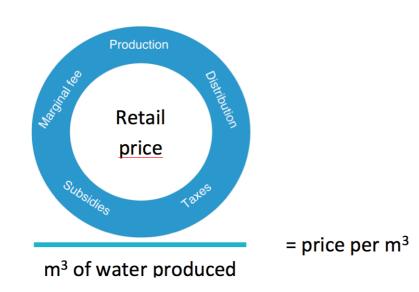
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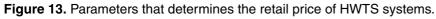


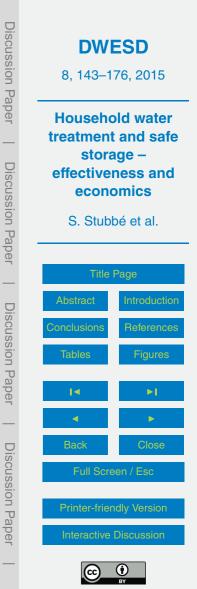


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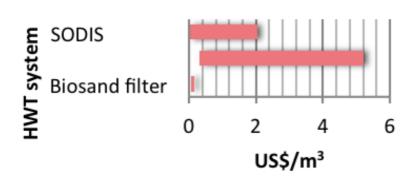


Figure 14. Price ranges (USDm⁻³) for the three selected HWTS systems (outliner of biosand filter is neglected).

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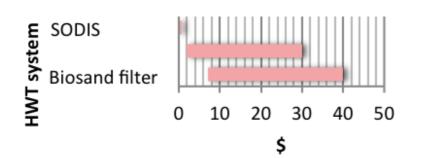


Figure 15. Ranges for the retail price of the three selected HWTS systems (outliner of SODIS is neglected).

